## State of Bird Populations in Britain and Ireland

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#### Abstract

The bird fauna of Britain and Ireland has been studied for over 500 years and changes in numbers over the last 40 years are well documented by a range of surveys, mostly undertaken by volunteers. Britain and Ireland are home to internationally important numbers of seabirds in the breeding season and many species of wildfowl and waders in the winter. Although there has been little overall change in total bird numbers, about half of bird species in Britain and Ireland are of conservation concern because of small or declining populations. In the last 30 years declines in farmland birds have been well documented, but declines in many woodland and migratory birds are just beginning to be recognised. Conversely, there have been increases in numbers of many raptors, seabirds and waterbirds. In general, habitat specialists have tended to decline most, while more adaptable, generalist species are increasing in number. Climate change threatens bird populations in all habitats, but loss of habitat and deterioration in habitat quality are more pressing threats. Conservationists have been successful at increasing populations of rarer species but the greater challenge will be conserving birds in the wider countryside, which will require the integration of conservation goals with wider social and landscape policies.

# Introduction

More is known about the birds of Britain than the vertebrate fauna of virtually any other country. This knowledge stretches back at least five centuries, with William Turner the first to publish original observations in his *Avium praecipuarum quarum apud Plinium et Aristotlem mentio est brevis & succinta historia* [A short and succinct history of the principal birds noted by Pliny and Aristotle] in 1544 (Bircham 2007). Turner's 'succinct history' included just over 100 species that he recognized as occurring in Britain, including the phesan (pheasant<sup>1</sup>), bramlynge (brambling), nut-jobber (nuthatch) and solend guse (gannet). The first 'modern' ornithology was that of John Ray and Francis Willughby, which appeared in English (rather than Latin) in 1678 and which recognised around 200 species in Britain. By January 2009, 584 species had been recorded in Britain (and 460 in Ireland), representing an increase in knowledge and field effort as much as an increase in the number of species occurring (Parkin & Knox 2009).

The breeding population of birds in Britain is in the region of 75 million pairs of 220 species; a further 50 or so species visit regularly during the winter months or pass though in spring and autumn, though numbers of these are more poorly known (Appendix 1). About 200 species occur regularly in Ireland, reflecting its position further from the European mainland. Overall, populations of Britain's commoner breeding birds seem

<sup>&</sup>lt;sup>1</sup> A note on names: common names follow those in widest use. Any ambiguity should be resolved by reference to scientific names, which are given in the text at first mention or in Appendix 1.

remarkably stable (Fig. 1). This simple line, though, hides much variation, with some species (e.g. seabirds) increasing in numbers, whilst others (e.g. farmland birds) have fallen dramatically. The simplicity of this figure also belies the vast amount of effort, almost all unpaid, undertaken in survey work to produce the data on which it is based. However, its strength lies in its very simplicity: the clear decline in farmland birds it demonstrates, for example, was instrumental in stimulating a policy response to address the issues underlying the decline.

In comparison to other taxa, birds are relatively to easy to monitor being popular, visible, (mostly) diurnal and (generally) countable; consequently a large amount is known about their trends, at least in recent decades and for the commoner species (Appendix 1). Rather than pick out individual species, I have tried to identify common patterns that reflect the major patterns of environmental change in our islands over the last 40 years before going on to outline some of the major challenges for bird conservation in the coming decades. But first I would like to give a flavour of the range of monitoring efforts that exist which mean we know so much about the state of our avian fauna.

# Monitoring bird populations: approaches and scope

Birds vary greatly in number: from a single long-billed murrelet *Brachyramphus perdix* recorded off the Devon coast in 2006 to the wren, which may, in some years, number tens of millions of individuals. They also vary greatly in their ecology, from nightingales inhabiting dense scrub of southern Britain to the ptarmigan of the windswept Scottish uplands, and in how easy they are count, from the gannet which nests mostly on offshore stacks in Ireland and Scotland and whose population can be almost completely enumerated using aerial photography, to the cryptic and largely nocturnal woodcock for which population estimates are little more than educated guesses. Surveys of abundance therefore differ greatly in their aims and methods.

For some species, typically those with localised distributions, a complete, or almost complete, census can be contemplated; for commoner and widespread species, sample surveys must be undertaken. Large-scale sample surveys provide detailed information on an annual basis, while for trends in species monitored by census are known in less detail as these are usually less than annual owing to the greater effort required. The annual *State of the UK's birds* provides summary results from most monitoring schemes each year (e.g. Eaton et al. 2008). For rarer species collation of reported sightings may be all that is possible (Table 1). : annual summaries are produced by the Rare Breeding Birds Panel (most recently, Holling & RBBP 2008), the *British Birds* Rarities Committee (Hudson & BBRC 2008) and their Irish equivalents (Hillis 2008; Milne & McAdams 2008), as is a summary of the numbers of scarce migrants reported each year (Rogers 2006; Rattigan 2007).

Most monitoring of birds is undertaken by unpaid volunteers working in conjunction with professional scientists who direct the sampling and analyse the data collected to inform conservation and land management policies (Fig. 2). This arrangement not only allows schemes to operate over the long-term and extensive spatial scales at very low cost, it

also encourages a greater appreciation of science and the environment by participants, facilitating democratic participation in research that often has a direct impact on wider policy (Greenwood 2007).

Britain and Ireland have a long tradition of recording birds stretching back into the 19<sup>th</sup> Century; though this information is rarely quantitative, broad patterns of distribution can be mapped (Holloway 1996). Concerns over the use of pesticides in the 1950s, similar to those that led Rachel Carson to pen *Silent Spring*, prompted the Nature Conservancy (as it was then) to fund the British Trust for Ornithology (BTO) to organise an annual survey (the Common Birds Census, CBC) to measure both background variation in bird numbers and the extent of any changes as a result of pesticide poisoning, pollution or habitat change (Marchant et al. 1990). More recently this has been replaced by the Breeding Bird Survey (BBS), which provides more representative coverage of a greater range of species nationally, but less detailed information at each site (Newson et al. 2008).

For species that are too scarce to occur on a sufficient number of CBC or BBS sites, targeted surveys must be organised, often using specifically tailored methods. Species covered in this manner range from seabirds through Dartford warbler to golden eagle, with an honourable mention to the Heronries Census which began in 1928 and is the longest running annual single species survey in the world. Summary results from these surveys are presented in the annual *State of the UK's Birds* report (published jointly by Britain's bird conservation organisations) and at www.bto.org/birdtrends. These specific surveys are supplemented, every 20 years or so, by national atlases which aim to record the distribution (and relative abundance) of all regularly occurring species (Marchant et al. 2004). Most counties in Britain have also produced at least one local atlas (Ballance 2000) and the next atlas covering Britain & Ireland will be published in 2013 (www.birdatlas.net).

Monitoring of numbers (or relative abundance) is really just the first step. There is little point in monitoring unless appropriate action follows changes (up or down); determining the threshold at which action is required is a key conservation priority. Long-term quantitative data on trends mean that the status of species can be objectively assessed, resulting in much greater transparency in the conservation designation process (Eaton et al. 2009). Amongst waterbirds, high quality counts have also allowed site designation criteria (e.g. of Specially Protected Areas) to be based on the number of birds using the site as a proportion of the flyway population. They also enable alerts to be issued when populations decrease at site, regional or national level, helping to direct conservation and management policies (Atkinson et al. 2006). Understanding the cause of population change is aided by knowledge of the demography of the population, i.e. reproductive output, survival and dispersal between populations (Fig. 2). Combining information on population size, demography and environmental processes can yield powerful insights into the mechanisms of population change (e.g. Frederiksen et al. 2004; Robinson et al. 2004; Freeman et al. 2007b). This can be critical in identifying the key environmental drivers of change and effective management actions (Baillie 1990). Such actions may be required both if species are declining, of if they increase to the extent where they cause economic impacts on other users of the countryside.

## The changing state of Britain and Ireland's birds

Because they are islands, Britain and, especially, Ireland have smaller bird faunas than might be expected, with several species that are common on the near continent missing (Fuller et al. 2007). Some species are also at the edge of their range, either northern, such as reed warbler and nuthatch, or southern, such as dotterel and snow bunting. Their oceanic position and extensive coastline mean our islands support large populations of breeding seabirds; more than 20% of the European population of nine species, and more than half of the world population of three: gannet, bonxie (great skua) and Manx shearwater breed in Britain or Ireland (Burfield & van Bommel 2004). In winter, Britain and Ireland also support internationally important populations of many geese and waders (Delany & Scott 2006; Appendix 1). Only two species of global conservation concern occur regularly: Balearic shearwater (Critically Endangered) and aquatic warbler (Vulnerable), both on passage. A further six are classified as Near Threatened: sooty shearwater, red kite, corncrake, curlew, black-tailed godwit and Dartford warbler. Only one species, Scottish crossbill, is endemic but around 35 species have endemic races, four of which are restricted to Ireland.

One species occurring regularly in Britain and Ireland has become globally extinct: the last great auk *Pinguinis impennis* in Britain, was killed on Stac an Armin, St Kilda in the 1840's on the tragic presumption that it was a witch. Several species, though, have essentially disappeared as breeding birds in these islands but remain common elsewhere including: great bustard *Otis tarda* (extinct by mid 19<sup>th</sup> Century), black tern and Kentish plover *Charadrius alexandrinus* (mid 20<sup>th</sup> Century), wryneck, red-backed shrike and, in Ireland, corn bunting (late 20<sup>th</sup> Century). In the last 50 years though, these have been offset by re-colonising former breeders, such as avocet new breeding species, notably little egret, collared dove and, in England, Cetti's warbler, and introduced, non-native, species such as Egyptian goose (Fig. 3) and rose-ringed parakeet.

Comprehensive accounts of changes in individual species' status are available in Brown & Grice (2005, England), Forrester & Andrews (2007, Scotland), Hutchinson (1989, Ireland) and Lovegrove et al. (1994, Wales) and summarised in Parkin & Knox (2010). Because man has had such a huge influence on the British and Irish landscape, bird populations in particular habitats often exhibit similar trends in response to common environmental drivers (Fuller & Ausden 2008). I have structured the text around habitats to identify common themes among species inhabiting them. Most of the patterns described are common to both Britain and Ireland, as many of the drivers and trends are similar in these two countries; however data are much more readily available for Britain, which has had a stronger, and longer, history of monitoring bird populations, so these are more often quoted.

#### Marine

Seabirds are mostly ground and burrow nesters and consequently tend to nest on remote cliffs and islands where nest predators are few. Nowhere has this been more evident than

in the west of Scotland, where breeding seabirds, such as tystie (black guillemot) and Arctic tern, have all but disappeared from many locations as a result of depredation by American mink *Mustela vison* (Craik 1997). More widely, ship rat *Rattus norvegicus* predation has been associated with local declines of Manx shearwater and puffin colonies. Eradication of both predators is underway at some colonies, but is an arduous task (e.g. Moore et al. 2003).

Food availability is a strong driver of seabird population dynamics. An increase in the availability of food, particularly discards from the fishing industry, over the last century has been associated with dramatic increases in the numbers of fulmar and bonxie (great skua) (Mitchell et al. 2004); cormorants, meanwhile, have prospered on well-stocked inland waters used by anglers (Carss & Ekins 2002). The importance of food supply for seabird populations has been demonstrated by complete breeding failures in some areas in recent years when food has been scarce of species, such as kittiwake and puffin, which rely on sandeels *Ammodytes* to feed their chicks. When sandeel stocks collapsed around Shetland between 1985 and 1990, very few chicks were reared successfully leading to population declines in several species. Although over-fishing played a part in this decline, most fishing activity occurs beyond the foraging range of colonies and often targets bigger sandeels than seabirds forage on. Increases in sea temperatures forcing the sandeel spawning stock northwards, away from the colonies, seem to have been more important (Frederiksen et al. 2004).

Seabirds spend most of their lives at sea, only coming ashore to breed for two to three months each year. Despite the importance of offshore areas, relatively little is known about the status of seabirds at sea, although newly developed technologies, such as geolocators and data-loggers are providing exciting insights into the use of marine habitats far from shore. During the breeding season parents probably forage in hotspots near thermal fronts or areas of upwelling which tend to be rich in food, but in winter birds are much more dispersed (Stone et al. 1995). In recent years oil spills have killed many thousands of birds, particularly guillemots and other auks, but this seems to have had rather little impact on breeding populations due to the presence of a surplus of immature and non-breeding birds (Votier et al. 2008).

# Coasts and estuaries

Populations of many wintering ducks and geese was limited by hunting in the first half of the  $20^{\text{th}}$  century (Tubbs 1996). However, increased protection, particularly following the 1954 Protection of Birds Act, and a decline in the popularity of wildfowling have meant that numbers of most species have increased. Some have grown dramatically, numbers of pink-footed goose, rose from 30,000 in the early 1950s to around 250,000 today, for example. Similarly, most wintering estuarine wader populations are stable or increasing, for example, grey plover and black-tailed godwit (the Icelandic race *L. limosa islandica*); this reflects both a decrease in hunting pressure and, for at least some species, habitat changes on the breeding grounds (e.g. Gunnarsson et al. 2005). Numbers of birds on non-estuarine coastal sites (where there was less of a hunting tradition), however, have shown rather more mixed trends (Rehfisch et al. 2003) with some species typical of rocky

shores, such as purple sandpiper, turnstone and ringed plover being recorded in smaller numbers more recently.

Habitat on the coastal fringe has been under threat since large-scale drainage of coastal marshes began in the late 17<sup>th</sup> century. Currently, the major threat comes from industrial development, such as new ports, renewable energy generation (offshore wind and tidal stream) and reclamation for amenity and other uses. Although populations of many waders are continuing to increase, probably from artificially low populations created by hunting pressure, this continued loss of habitat must limit their capacity for increase. In the future, coastal habitat will also be increasingly squeezed between rising sea-levels, as a result of climate change, and sea-defences, particularly in south-east England.

(Atkinson et al. 2004). Climate change is already affecting shorebird populations. The Atlantic coasts of Britain and Ireland provide a mild climate, but good foraging habitat (invertebrate rich mudflats) is scarce. Conversely, our eastern shores have large estuaries but winters can be much more severe with prolonged spells of cold weather resulting in large-scale mortality. In recent years, there has been a tendency for winters in Europe to become warmer and, consequently, we are seeing a shift in winter distribution eastward as juveniles, particularly of Fennoscandian and Siberian breeding populations, settle eastward (Maclean et al. 2008). This reflects an increased carrying capacity of eastern estuaries allowing increased juvenile settlement on estuaries closer to their breeding grounds. Adult birds show high site-fidelity so are probably showing little change in distribution does not pose a threat to populations providing sufficient habitat is present, but it does mean that site designations (usually based on numbers using a site) may need to be reassessed and that conservation strategies will increasingly need an international dimension, considering the flyway in its entirety.

# Farmland

One of the most obvious changes in land-use over the last fifty years has been the intensification of agriculture and the loss of mixed farming (Chamberlain et al. 2000; Robinson & Sutherland 2002). Populations of farmland bird species mirror changes in agriculture throughout the 20<sup>th</sup> Century (Shrubb 2003). The area of agricultural land declined markedly between the 1870s and 1930s, but the post-war desire for food selfsufficiency saw a boom in cultivation and populations of many farmland species increased greatly in numbers, probably becoming more common in the 1970s than before or since. From around the point of Britain's entry into the European Community, advances in technology and generous production subsidies, such as guaranteed prices, meant that practices rapidly began to change: new crop varieties were introduced, crop rotations changed, chemical inputs increased, and farms became ever larger, specialising in either arable crops or livestock because of the need for expensive machinery (Fig. 4). Changes in pastoral farming have been equally marked with increased fertiliser usage, a switch from hay to silage cropping and increased stocking densities (Vickery et al. 2001). The primary effect of these changes has been a decrease in habitat heterogeneity at all scales, from within fields to across landscapes (Benton et al. 2003).

These changes have had a catastrophic effect on farmland bird populations (Fig. 1), with a litany of declines in the last 40 years amongst the species most closely associated with farmed land: tree sparrow 97%, corn bunting 86%, grey partridge 87% and skylark  $59\%^2$ . Declines in many once common species such as quail, corncrake and cirl bunting preceded these, but went unquantified (Fuller 2000; Newton 2004). Of course, some species increased, notably woodpigeon *Columba palumbus* (143%) and corvids, such as Jackdaw *Corvus monedula* (87%), probably as a result of reduced persecution and, in the case of woodpigeon, extensive sowing of oilseed rape. Although the general intensification of agriculture has undoubtedly been responsible for most of these changes, disentangling the impacts of and relative importance of the different facets of agricultural change is difficult. Doing so is, however, important for designing effective conservation solutions and in the last 20 years has been a major research area that has fed directly into, and been informed by, policy (e.g. Vickery et al. 2004). The long history of research on grey partridge provides an exemplary study in unravelling the causes of these declines (see Chapter 18).

Targeted conservation measures have proved easier to implement for species with restricted range than for more widespread species. So, for example, management agreements with farmers in south-west England have increased the cirl bunting population four-fold since 1989 in south-west England and slowed the decline of corncrakes (Jeffs & Evans 2004) and similar action has stabilised numbers of corncrake in Ireland (Copland 2002). Although agri-environment schemes can influence breeding populations (Gillings et al. 2005), their success has been mixed (Kleijn & Sutherland 2003; Ausden & Fuller 2009) and whether the recent stabilisation of farmland bird populations (Fig. 1) is due to changes in policy, a recent run of mild winters, or because populations are simply reaching a new carrying capacity remains to be seen. The future of farmland bird populations will continue to depend on the economics of agriculture. Increases in cereal prices, for example, resulted in the abolition of set-aside for 2008, greatly reducing the availability of winter foraging habitat for many species. Ongoing monitoring, though, means the effects of such changes can be quantified and, because of this, biodiversity targets are increasingly being incorporated alongside production in framing agricultural policies.

# Wetlands

A major component of agricultural intensification has been land drainage. Much lowland wet grassland has been lost, as well as unique habitats such as reed and sedge fens in the south and east of Britain and the seasonally flooded callows and turloughs of Ireland (Wakeham-Dawson & Smith 2000). Consequently, some fen specialists, such as black tern and bittern have all but disappeared as breeding birds and species associated with lowland wet grassland, notably breeding waders, such as lapwing, snipe and redshank, have declined significantly (Wilson *et al.* 2005). In addition to direct habitat loss, wetlands are vulnerable to increased nutrient loads (eutrophication), mostly from non-

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point sources such as leaching from agricultural land, and scrub encroachment, particularly as a result of lowered water tables, often reducing habitat quality for those that remain.

The expansion of cities and towns in the south of England and elsewhere required the extraction of large amounts of sand and gravel for construction, holes that were often turned into reservoirs to supply water, such as those at Abberton, Chew and Rutland. Many of these newly created water bodies have since been managed at least partly with nature conservation in mind, both for species conservation and general amenity value. This has provided large amounts of new habitat for breeding and wintering waterbirds, such as great-crested grebe, tufted duck and gadwall, and inland breeding waders, notably little ringed plover.

In the last twenty years much effort has been expended in re-creating wetlands in areas which have suffered large historic losses, such as the Thames estuary and the fens of East Anglia (Hume 2008, see also www.wetlandvision.org.uk). Such projects are often targeted at 'flagship' species, particularly bittern, but many other species (and not just birds) are also benefiting. Such habitat creation is likely to become commoner in the future as developers are increasingly required to compensate for lost habitat. Habitat creation schemes seem to be more successful for wetland (and coastal) habitats than terrestrial ones (Morris et al. 2006), though whether entire communities can be simply be translocated and re-created elsewhere is a question that remains unanswered, both ecologically and morally.

# Uplands

Despite their apparent bleakness, upland habitats, which range from unenclosed grazed pasture through bog and plantation forestry to arctic-alpine tundra on the high tops, support a diverse bird fauna including many of our most emblematic birds, such as golden eagle, raven and red grouse (Ratcliffe 1990). Open moor is itself the result of a long-history of land-use extending back nearly 4,000 years in some regions; much of what might once have been montane scrub and wood is now sheepwalk or acid grassland. Britain, particularly Scotland, and Ireland are internationally important for blanket and raised bogs, despite the fact that less than 20% of their original area remains. Bogs have been encroached by drainage for agriculture, afforestation and, especially in Ireland, large-scale extraction of peat for fuel and horticulture. Consequently, numbers of birds of open moor and bog, such as curlew (Fig. 5), dunlin and red grouse, have declined through habitat loss (Sim et al. 2005, Pearce-Higgins et al. 2008).

Although the uplands have long been grazed, they cannot withstand the current level of grazing pressure. High densities of red deer *Cervus elaphus* and sheep have greatly reduced vegetation diversity, diminishing both nesting and foraging habitat for birds (Fuller & Gough 1999; Clutton-Brock 2004). The number of sheep in upland areas is beginning to decrease with a shift away from headage payments, leading to scrub encroachment in some areas, which will benefit some species, such as willow warbler; deer numbers, on the other hand, are likely to continue to increase. Agriculture is also

encroaching on the uplands with marginal land being enclosed and improved (fertilised) for grazing, higher levels of drainage and an increase in silage cropping. These changes appear to have precipitated declines in a number of species associated with upland grassland, such as whinchat, wheatear and ring ousel (Henderson et al. 2004). As in lowland areas, the fate of upland bird populations will depend on the degree of habitat diversity that can be maintained at different scales (Pearce-Higgins & Grant 2006).

Although populations of moorland birds have suffered from loss of habitat due to afforestation, upland forests can provide important habitat for some species, particularly in the first 10-15 years before the canopy closes (Avery & Leslie 1990). The spread of young forest aided the recolonisation of mainland Scotland by hen harriers from the Northern Isles in the 1940s, for example. Middle-aged plantations, with a dense, dark canopy support relatively few species, but as the plantation matures, other species, such as crossbill, can colonise, so one bird community is successively replaced by another. Consequently, as plantations have matured and the area of new planting fallen, we have seen declines in tree pipit and lesser redpoll but increases in goshawk and siskin populations. Much of the conservation value of moorland communities, however, lies in their distinctness and rarity in a European context, while most species typical of conifer plantations are common throughout Europe, so increased diversity may not be a sufficient goal in itself. A key issue for the Scottish uplands is the extent to which we wish to encourage natural regeneration of broadleaf and Scots pine Pinus sylvestris woodland (replacing the presumed 'wood of Caledon'), which again will benefit some bird species, but be detrimental to others.

Birds of the high tops, such as ptarmigan, dotterel and snow bunting, are at the southern edge of their range in Britain and prospects for their continued existence in Britain look bleak, as the high tops are amongst the areas most likely to be affected by increasing global temperatures. In the past few decades, for example, the area of lying snow has decreased markedly and, although there is little hard evidence, it seems likely that the area suitable for breeding for most of these species will decline over the next two or three decades. Whether management actions can be identified to preserve these populations, which are also suffering from increased visitor pressure, remains to be seen.

While birds of prey were once widespread across Britain, ranges of many contracted into upland areas, away from human populations, as a result of persecution and poisoning from pesticides, particularly organochlorines such as dieldrin (Newton 1979). The uplands still represent strongholds for many species, but the phasing out of organochlorine pesticides as their environmental impacts became clear and increased legal protection afforded by the 1954 Protection of Birds Act mean that most species have increased in number and expanded back into lowland areas (Greenwood et al. 2003). Persecution of raptors has often been greatest in areas of heather *Calluna* moor managed for driven grouse shooting. Relationships between raptor numbers, grouse production and habitat are complex and reductions in management intensity and overgrazing are leading to a general decline in the quality of grouse moors, both in economic and biodiversity terms (Thirgood et al. 2000). Despite increased protection,

continued persecution in some areas appears to be preventing further recovery of hen harriers and other raptors.

## Woodland & Scrub

Although much of lowland Britain and Ireland would once have been woodland (not necessarily with a continuous canopy), most of our woods have for a long time been relatively fragmented and heavily modified by man (Fuller 1995). The area of seminatural woodland (and particularly ancient woodland) is gradually reducing, but the total area of woodland has steadily increased over the last forty years, mostly through commercial planting of conifers (Mason 2007). Population trends among woodland bird species have not been consistent over the last 40 years: scarce woodland specialists, such as the lesser spotted woodpecker and willow tit, and long-distance migrants, such as spotted flycatcher and tree pipit, tend to be declining while more generalist species tend to be increasing (Fuller et al. 2005).

Undoubtedly, greatly increased deer numbers, particularly of roe deer *Capreolus capreolus* and muntjac *Muntiacus reevesi*, are having a major impact on our woodlands, resulting in a reduction in habitat quality and changes in the structure of the woodland understorey which many birds nest or forage in (Fuller 2001). In the last two to three decades there has also been a notable decrease in the intensity of woodland management; denser canopies have shaded out shrub layer plants, such as bramble *Rubus fruticosus*, to the detriment of species like nightingale, which prefer a dense shrub layer (Fuller et al. 2005). It is possible that loss of habitat diversity as a result of deer browsing and changed stand management may also increase levels of nest predation (Evans 2004). Looking to the future, short-rotation coppicing, particularly of willow *Salix*, holds potential as a source of biofuel, but its impact on bird communities will depend on where it is planted and how it is managed (Anderson & Fergusson 2006).

Scrub is a bit of a Cinderella habitat, always on the boundary. Although there are places where scrub is probably the natural habitat, storm-lashed coastal heaths and in submontane areas above the tree-line, it is more often a (usually unwelcome) stage in the succession of vegetation from open heath or marsh to young woodland. Yet scrub supports a distinctive community of birds, albeit one which is sensitive to the amount of cover available with, for example, tree pipit and linnet on more open heaths and garden warbler and dunnock commoner in areas with continuous canopy. They also provide an important source of food for berry-eating thrushes and others in winter. The fortunes of such species undoubtedly reflect the amount of available habitat and many will have declined due to loss of habitat to agriculture or woodland succession. More recently active habitat management has seen a great improvement in the overall condition of Britain's shrub and heath stock, so many species, like the Dartford warbler and nightjar, are currently prospering.

Afro-Palaearctic migrants are an important component of woodland and scrub communities and many species, particularly those that migrate to central and southern Africa, are declining (Hewson et al. 2007). Similar declines have been seen amongst

long-distance migrants in other habitats, and across Europe (Sanderson et al. 2006). The reasons for these declines are unclear, but could include: deterioration in habitat quality (either in Europe or Africa), increased competition from more abundant resident species as a result of milder winters, or changes in seasonal phenology. Many bird species are breeding earlier in response to warmer springs but so are their insect prey, which are emerging and maturing earlier (Crick 2004). Migrants may be constrained in the degree to which they can advance their timing of breeding because of the need to migrate back to breeding grounds, whereas insects can respond to warming temperatures more rapidly and to a greater extent. There is some evidence that in areas where insect phenology has advanced to the greatest extent, populations of pied flycatchers are doing least well because their timing of breeding is no longer so closely matched with their prey (*e.g.* Both et al. 2006). Though the extent to which these findings generalise to other regions and species is unknown, such decoupling has the potential to profoundly alter ecological systems in a way that will be difficult mitigate.

## Town and gardens

Birds of town and garden are often over-looked in inventories of biodiversity, yet the area of private gardens is estimated to be approximately twice that of nature reserves (local, national and RSPB) in England & Wales (Cannon et al. 2005). Gardens support a significant proportion of the national population for some species (Gregory & Baillie 1998); the BBS, for example, identifies eight species as characteristic of urban and suburban areas: house sparrow (Fig. 6), starling, blackbird, magpie, collared dove, greenfinch, carrion crow and, increasingly, woodpigeon. They can also provide an important refuge for some species of conservation concern, such as song thrush and spotted flycatcher (Bland et al. 2004).

Birds of gardens tend to be those that are common (or, well, garden!) and populations of such species are generally increasing, at least in part because of the increased provision (and in greater variety) of food (Jones & Reynolds 2008). Food provision can also be important for species in the wider countryside, for example, goldfinches are increasingly exploiting garden feeders in winter as seed supplies in farmland decrease. Urban areas can also provide other resources, for example, the number of herring gulls nesting on flat roofs is increasing markedly, in contrast to declining populations in coastal areas (Raven & Coulson 1997). However, even within towns, though, suitable habitat can be quite patchy, for example, house sparrows, once ubiquitous, now tend to be associated only with particular types of housing (Shaw et al. 2008) for reasons which are unclear but which may be related to the availability of nest sites or foraging opportunities.

Urban expansion still represents a threat of habitat loss and fragmentation. Over the next twenty years, around 200,000 new homes are expected to be required annually in England and there is a limit to how many can be built on brownfield sites (Barker 2006). Such effects extend beyond direct habitat loss, for example the number of breeding nightjars is lower on heathland closer to towns because of greater disturbance from visitors (Liley & Clarke 2003). Whether human disturbance actually affects population

levels is unknown, but access to, and management of, the countryside will continue to increase, so quantifying such impacts will be important (Sutherland 2007).

## Summary

Overall, more species have increased than decreased over the last three to four decades (Fig. 7), however, about half of Britain's and Ireland's breeding species are of conservation concern in a national context because of small, declining or concentrated populations (Gregory et al. 2002; Lynas et al. 2007). Declines amongst farmland birds have been well documented, but widespread declines in woodland and migratory species are only just being recognised. Even populations of species once considered as pests, such as house sparrow and bullfinch, have declined by as much as 50%. More generally, habitat specialists have tended to decline while commoner, more adaptable species have increased. Habitat specialists tend to be those of most conservation interest precisely because they are often restricted in range or abundance. Historically, habitat loss and fragmentation have been major drivers of population change, but deterioration in habitat quality is now at least as important (Fuller & Ausden 2008).

Rarer species (those with fewer than 1,000 individuals) have tended to fare better, while scarce species (those with fewer than 100,000 individuals) are doing worse (Fig. 8). Targeting rarer species, for example cirl bunting (agreements with farmers) and goldeneye (providing nest boxes), with conservation measures is relatively easy; increasing numbers of more widespread species that are currently declining will be much harder. Many species that suffered from persecution have benefited from increased legal protection, so most raptors, for instance, are increasing (although not in all areas). Many wildfowl populations are also increasing, though coastal habitats continue to be severely threatened both from human development and a changing climate. Similarly, although seabird populations have increased in recent decades, whether this will hold true as climates change remains to be seen.

# The future for Britain and Ireland's birds

Bird conservation has a long history in Britain: St Cuthbert created a sanctuary for birds on Inner Farne (Northumberland) in 676; the first Parliamentary Act (protecting game species) was passed in 1831 and the Society for the Protection of Birds was formed in 1889 (it gained its Royal Charter in 1904). An idea of the popularity (and influence) of bird conservation can be gained from membership of the RSPB, which currently stands at just over one million, nearly twice that of the combined membership of Britain's three main political parties (c. 560,000). Reflecting this interest and the availability of data, the UK government has adopted an index of bird numbers as a surrogate for biodiversity generally in its measure of environmental sustainability, an initiative subsequently followed by the European Union (Gregory et al. 2005). Despite their flaws, such simple measures can be highly effective in communicating the need for action to policy makers and others. High-profile re-introduction projects, such as those of red kite and whitetailed eagle, have also been influential in engaging interest in conservation, as well as providing income for local economies (Carter et al. 2007. People have a greater interest in bird conservation if they see its benefits directly, and politicians need to know it matters to their electorate, as well as what solutions are available to particular problems).

Bird conservation is increasingly moving beyond the nature reserve. It is no longer just about protecting particular species on nature reserves (though this will, of course, remain an important challenge), but rather the much more difficult task of integrating conservation goals with other landscape planning priorities (Sutherland 2004; Robinson 2006). For example, wild geese often forage on farmland creating an economic conflict between conservationists and farmers, one that has increased as goose populations have grown (Vickery & Gill 1999). More generally, is farmland solely for producing food, or should it also provide habitats for wildlife? More generally, should we have a smaller area of intensively farmed land, leaving room for undisturbed semi-natural habitat elsewhere, or more extensive areas of less-intensively managed land (e.g. Green et al. 2005)? The answers to such questions need to recognise that people value, and gain benefits from, the biodiversity immediately around them (Pretty et al. 2007). This often means birds as the most visible component of our fauna, but the balance between food production and biodiversity is one for society to determine.

The drivers of bird population change increasingly act at an international scale; trends in bird populations in Britain are similar to those in (north-western) continental Europe (Burfield & van Bommel 2004). For example, the loss of farmland birds in Britain is mirrored across western Europe reflecting common agricultural policies (Donald et al. 2001). Trends in migratory birds are also similar across Europe, perhaps reflecting changes in shared wintering quarters, similar patterns of land-use across Europe, or a changing global climate (Sanderson et al. 2006). Bird conservation thus needs to be coordinated internationally and there is evidence that a coherent site protection policy across countries, such as the Natura 2000 network, can deliver increased bird populations (Donald et al. 2007).

In the longer-term, changes in climate will have an over-arching impact on bird populations, either directly, for example, rising sea-levels reducing coastal habitat (Crick 2004), or indirectly, perhaps by altering patterns of agriculture or development (Olesen & Bindi 2002). The most demonstrable effect of climate change on bird populations has been changes of range: the northern edge of bird distributions is moving north, both during the summer (on average by 20km between 1970 and 1990, Thomas & Lennon 1999) and winter (e.g. wader distributions have moved northeast by 85km since the late 1970s, Maclean et al. 2008) and these changes are predicted to continue (perhaps by as much as 550km across Europe by the end of the century, Huntley et al. 2008). Such changes are clearly profound, and may result in new avian communities evolving with complex, and unpredictable, consequences. Huntley et al. (2008) predict future ranges will only overlap by 40% with current ranges, so a key priority will be ensuring the landscape is sufficiently permeable to allow ranges to shift and that suitable habitat is available to move into. This will be a challenging goal given the current intensity of land-use over much of Europe.

In Britain, the main legislation protecting birds are the Wildlife and Countryside (1981, and amendments) and the Countryside and Rights of Way (2000) Acts. Increasingly, though, there is a need for internationally co-ordinated legislation such as the European Birds Directive (79/409/EEC), which provides for Special Protection Areas (SPAs) to protect bird populations which, along with Special Areas of Conservation (SACs) designated under the Habitats Directive (92/43/EEC), form the core of the Natura 2000 reserve network across Europe. This network has increased numbers of species listed under the Directive in countries where it has been developed (Donald et al. 2007). With changing climates such co-ordinated networks of protected sites will increasingly be needed if species are to adapt their ranges appropriately.

Bird populations have changed dramatically over the last 15,000 years (Yalden & Albarella 2008) and even over the last 200 years (Gibbons et al. 1996); future changes, both positive and negative, are inevitable and will be contingent on our management of the landscape. We cannot preserve our countryside or wildlife in aspic, so what is needed is a coherent vision of the mix of habitats we would like to see; that is moving away from conserving 'original-natural' landscapes, which may be semi-mythical anyway, to creating 'future-natural' landscapes which support functioning ecosystems (Ausden & Fuller 2009). These will be expensive to create, so biodiversity goals need to be integrated with ecosystem services (such as carbon sequestration or flood management) to satisfy wider social and land management aims. In broad terms, we probably have sufficient understanding to benefit populations of scarce specialists, especially within protected areas, and recent increases in stone curlew, corncrake and cirl bunting, amongst others, show how effective evidence-based conservation measures can be (Aebischer et al. 2000). The greater challenge will be in maintaining landscapes that are sufficiently diverse that such intervention becomes unnecessary.

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Table 1. Monitoring of Britain and Ireland's bird populations. Key schemes currently operating are listed with an indication of their geographic scope and year of commencement and survey interval. Results are published in books (first author, year of publication and publisher are given), annual reports (with lead organisation) or as journal articles (most recent cited).

Breeding Sea	son			
All	Distribution Atlas	Brit. & Irel.	1972, c20yr	Sharrock (1976, Poyser); Gibbons et al. (1993, Poyser)
Common	Breeding Bird Survey	UK	1994, 1yr	www.bto.org/bbs
	Countryside Bird Survey (CBS)	Ireland	1998, 1 yr	Coombes et al. (2009), BirdWatch Ireland
	Waterway Breeding Bird Survey (WBBS)	UK	1998, 1 yr	www.bto.org/survey/wbbs.htm
Grey Heron	Heronries Census	UK	1928, 1 yr	Ibis 146:323-334
Seabirds	Seabird Colony Register	Brit. & Irel.	1969, c. 15yr	Cramp et al. (1974, Collins); Lloyd et al. (1991, Poyser); Mitchell et al. (2004, Poyser)
	Seabird Monitoring Programme (SMP)	UK	1986, 1 yr	www.jncc.gov.uk/page-1550
Scarce	Statutory Conservation Agency/RSPB Annual Breeding Bird Survey (SCARABBS)	UK	1961, various	
Rare	Rare Breeding Birds Panel (RBBP)	UK	1972, 1 yr	British Birds 103:2-52
	Irish Rare Breeding Birds Panel	Ireland	2002, 1 yr	Irish Birds 8:365-394
Non-native	Rare Breeding Birds Panel	UK	1996, 1yr	British Birds 100:638-649
Non-breeding	g Season			
All	Distribution Atlas	Brit. & Irel.	1983, c. 25 yr	Lack (1986, Poyser)
Seabirds	At-sea distribution	NW Europe		Stone et al. (1995, JNCC)
Waterbirds	Wetland Bird Survey (WeBS)	UK	1947, 1 yr	Annual report, BTO
	Irish Wetland Bird Survey (I-WeBS)	Ireland	1994, 1 yr	Irish Birds 8:341-350
	Non-estuarine Coastal Waterbird Survey (NeWS)	UK	1984, c. 10yr	Bird Study 50:22-32
Gulls	Winter Gull Roost Survey (WinGS)	UK	1953, 10 yr	British Birds 96:376-401
Scarce Migrants		Britain	1958, 1 yr	British Birds 99:74-117, 129-173
		Ireland	2004, 1 yr	Irish Birds 8:263-298
Rare	British Birds Rarities Committee (BBRC)	Britain	1958, 1 yr	British Birds 102:528-601
	Irish Rare Birds Committee (IRBC)	Ireland	1953, 1 yr	Irish Birds 8:395-416

Figure 1. Trends in UK bird populations 1970-2005. The thick line represents the average trend of 116 species of breeding bird species, and the thinner lines 19 species of farmland, 38 woodland and 20 seabird species. The index for wintering waterbirds (68 spp, dashed line) is not part of the All Species indicator. In each case the index (which is an unweighted average of the trends for each constituent species) is arbitrarily set to 1 in 1970 (1975 for waterbirds as earlier values are not available).

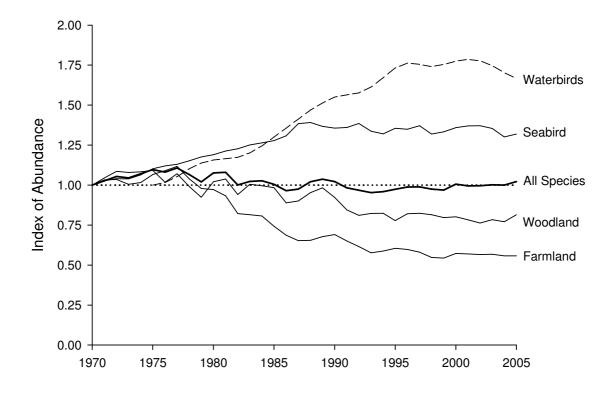
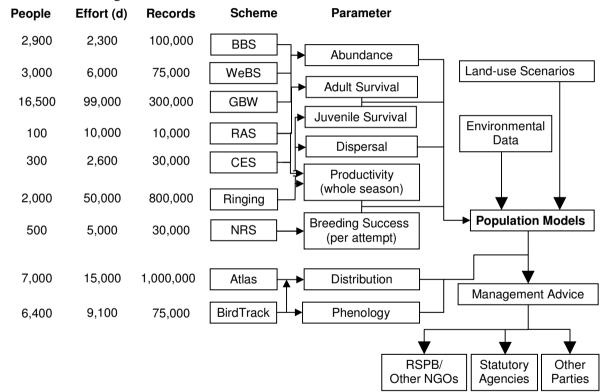


Figure 2. Integrated Population Monitoring. Each of the core monitoring schemes listed provides information on one or more key demographic parameters, which can be used to understand causes of population change through population modelling, which can then feed through into conservation or management advice to relevant stake holders. For each scheme number of participants, estimated time spent each year (days) and approximate number of records contributed annually is indicated. The schemes represented are a selection of the core annual monitoring schemes, several others also exist.



Footnote to Fig. 2. Schemes listed are: BBS –Breeding Bird Survey, GBW – BTO/CJ Garden BirdWatch, WeBS - Wetland Bird Survey, CES – Constant Effort Sites ringing scheme, RAS – Re-trapping Adults for Survival ringing scheme, NRS – Nest Record Scheme. For BBS, GBW, WeBS, Atlas (2007-2011) and BirdTrack a record is one species at one site, for NRS and the ringing schemes a record is an individual bird or nest.

Figure 4. Potential causes of population change in plant, insect and bird populations resulting from changes in arable management. The major drivers are highlighted by shading. From Robinson & Sutherland (2002)

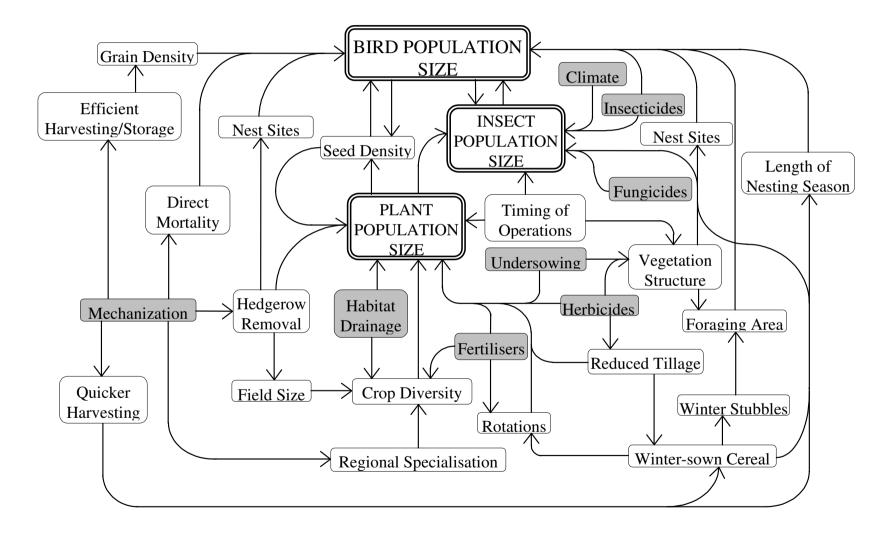


Figure 7. Population trends of British breeding birds in different habitats over the last 40 years. Species are indicated as strongly (black) or moderately (grey) increasing/decreasing or as stable/fluctuating (white). The dashed lines indicate the overall proportion of the 213 species that are increasing (38%) or decreasing (29%), also given is number of species included in each group. From data in Appendix 1.

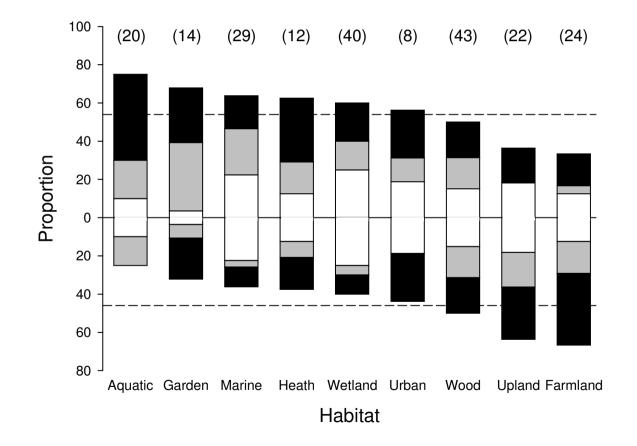
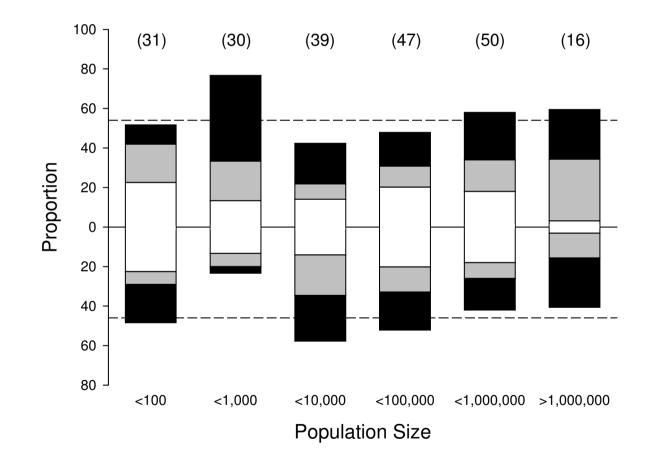


Figure 8. Population status of species breeding in Britain over the last 40 years. Species, grouped by population size, are indicated as strongly (black) or moderately (grey) increasing/decreasing or as stable or fluctuating (white). The dashed lines indicate the overall proportion of the 213 species that are increasing (38%) or decreasing (29%), also given is number of species included in each group. From data in Appendix 1.



**Appendix 1**. Population status of species regularly occurring in Britain. Conservation status (R[ed], A[mber] or G[reen], see Gregory et al. [2002]) and year of first successful breeding (after 1900) are given. Estimates of population size in 2000 (or the nearest available year) are taken from Baker et al. (2006), updated by Robinson (2005) and Newson *et al.* (2008); populations important (> 20% of total) in a European or flyway context are highlighted in bold, based on figures in Burfield & van Bommel (2004) and Delany & Scott (2006). A qualitative assessment of population trend since the 1960s is presented: extreme decline (---, >75%) or increase (+++, >100%), steep decline or increase (--/++, >50%), moderate decline or increase (-/+, >25%), or as stable or fluctuating (~, <25% change); if the trend differs substantially before and after the 1980s these are indicated separately. Where a source is given trends are based on more or less quantitative data, but the amount and quality of information varies hugely, so they are best regarded as indicative. Question marks indicate particular uncertainties.

Species	Status <sup>1</sup>		First Bred	Population <sup>2</sup>	Trend	Source <sup>3</sup>
Red-throated Diver Gavia stellata	RB WV	А		B: 1,200 P	+/~	
				W: 17,000 I	?	
Black-throated Diver Gavia arctica	RB WV	Α		B: 170 P	?+	
Great Northern Diver Gavia immer	CB WV	Α	1970	W: 2,800 I	?	
Little Grebe Tachybaptus ruficollis	RB WV	G		B: 7,500 P	?-	4
Great Crested Grebe Podiceps cristatus	RB WV	G		B: 23,000 I	++	6
Red-necked Grebe Podiceps grisegena	CB WV	Α	1988	W: 200 I	~	6
Slavonian Grebe Podiceps auritus	RB WV	Α	1908	B: 40 P	~/	10
Black-necked Grebe Podiceps nigricollis	RB WV	Α	1904	B: 50 P	+++	12
Fulmar Fulmarus glacialis	RB PV	Α	1878	B: 500,000 P	++/~	11
Cory's Shearwater Calonectris diomedea	PV	-		P: 700 I	?++	9
Great Shearwater Puffinus gravis	PV	G		P: >150 I	?	
Sooty Shearwater Puffinus griseus	PV	G		P: >1,000 I	?	
Manx Shearwater Puffinus puffinus	MB	Α		B: 300,000 P	?	11
Balearic Shearwater Puffinus mauretanicus	PV	-		P: >1,000 I	?+	
Storm-petrel Hydrobates pelagicus	MB	Α		B: 26,000 P	?	11
Leach's Storm-petrel Oceanodroma leucorhoa	MB PV	Α		B: 48,000 P	?	11
Gannet Morus bassanus	RB PV	Α		B: 220,000 N	++	11
Cormorant Phalacrocorax carbo	RB WV	Α		B: 8,400 P	+	11
				W: 23,000 I	+	6
Shag Phalacrocorax aristotelis	RB	Α		B: 27,000 P	~	11
Bittern Botaurus stellaris	RB WV	R	$1911^{+}$	B: 50 M	-/~	8
Cattle Egret Bubulcus ibis	CB SV	-	2008	P: 10 I	+	13
Little Egret Egretta garzetta	RB PV	Α	1995	B: 150 P	+++	8
Grey Heron Ardea cinerea	RB WV	G		B: 13,000 N	~/+	12
Purple Heron Ardea purpurea	PV	-	2010	P: 20 I	~	9
Spoonbill Platalea leucorodia	CB PV	Α	1998	P: 160 I	+	
Mute Swan Cygnus olor	RB	Α		B: 45,000 P	+++	4
Bewick's Swan Cygnus columbianus	WV	Α		W: 8,100 I	+++/~	6
Whooper Swan Cygnus cygnus	CB WV	Α	$1910^{\dagger}$	W: 5,700 I	~/+++	6
Bean Goose Anser fabalis	WV	Α		W: 500 I	~	6
Pink-footed Goose Anser brachyrhynchus	WV	Α		W: 240,000 I	+++	6
White-fronted Goose Anser albifrons	WV	Α		W: 27,000 I	++/-	6
* Greylag Goose Anser anser	RB WV	Α		B: 130,000 I	+++	4
				W: 120,000 I	++	6
<sup>*</sup> Canada Goose Branta canadensis	RB	-	1890	B: 82,000 I	+++	4
Barnacle Goose Branta leucopsis	WV	Α		W: 67,000 I	+++	6
Brent Goose Branta bernicla	WV	Α		W: 67,000 I	+++/~	6
* Egyptian Goose Alopochen aegyptiaca	RB	-	1700s	B: 1,000 I	+++	6
Shelduck Tadorna tadorna	RB WV	Α		B: 35,000 P	++/~	1
				W: 78,000 I	~	6

* Mandarin Duck Aix galericulata	RB	-	1928	B: 7,000 I	+++	
Wigeon Anas penelope	RB WV	А		B: 400 P	?	
				W: 410,000 I	++	6
Gadwall Anas strepera	RB WV	А	1850	B: 770 P	++	
				W: 17,000 I	+++	6
Teal Anas crecca	RB WV	А		B: 2,000 P	-	0
Teal Anus creecu		Α		W: 190,000 I	++	6
Malland Awar plate where the		C				2
Mallard Anas platyrhynchos	RB WV	G		B: 660,000 P	+++	
			10/0	W: 350,000 I		6
Pintail Anas acuta	RB WV	А	1869	B: 20 P	~ ,	8
				W: 28,000 I	+++/~	6
Garganey Anas querquedula	MB PV	А	1862	B: 70 P	~	
Shoveler Anas clypeata	MB WV	А		B: 1,200 P	+/-	
				W: 15,000 I	++	6
Pochard Aythya ferina	RB WV	Α		B: 460 P	+/~	8
				W: 60,000 I	-	6
Tufted Duck Aythya fuligula	RB WV	G	1849	B: 42,000 P	+	4
				W: 90,000 I	+	6
Scaup Aythya marila	CB WV	А	1897	W: 7,600 I	/~	6
Eider Somateria mollissima	RB WV	A		B: 31,000 P	?~	
				W: 73,000 I	++/~	6
Long-tailed Duck Clangula hyemalis	WV	А		W: 16,000 I	?-	0
Common Scoter <i>Melanitta nigra</i>	RB WV	R	1855	B: 100 P	?-	
Common Scoter Metanula nigra		К	1655	W: 50,000 I	?	
Malant Canton Malanitta Gana	11/17	•			?-	
Velvet Scoter Melanitta fusca	WV	A	1070	W: 3,000 I		
Goldeneye Bucephala clangula	RB WV	А	1970	B: 200 P	+++/~	
		~		W: 25,000 I	~	6
Smew Mergellus albellus	WV	G		W: 400 I	~	6
Red-breasted Merganser Mergus serrator	RB WV	G		B: 2,200 P	?~	12
				W: 9,800 I	++/~	6
Goosander Mergus merganser	RB WV	G	1871	B: 2,600 P	+++	4
				W: 16,000 I	+	6
* Ruddy Duck Oxyura jamaicensis	RB	-	1960	B: <600 P	+++	6
Honey-buzzard Pernis apivorus	MB PV	А		B: 50 P	~/+++	8
* Red Kite Milvus milvus	RB WV	А		B: 1,000 P	+++	8
*White-tailed Eagle Haliaeetus albicilla	RB	R	1983 <sup>†</sup>	B: 40 P	+++	8
Marsh Harrier Circus aeruginosus	MB PV	А		B: 360 F	+++	8
Hen Harrier <i>Circus cyaneus</i>	RB WV	R		B: 800 P	~	12
Montagu's Harrier <i>Circus pygargus</i>	MB	A		B: 10 T	+	8
Goshawk Accipiter gentilis	RB	G	$1938^{\dagger}$	B: 400 P	+++	8
Sparrowhawk Accipiter nisus	RB WV	G	1750	B: 40,000 P	+++/~	2
1 · · · ·	RB	G		B: 40,000 F B: 38,000 T	+++/~ ~/+++	$\frac{2}{2}$
Buzzard Buteo buteo		G				
Rough-legged Buzzard Buteo lagopus	WV	-		W: 40 I	?+	9
Golden Eagle Aquila chrysaetos	RB	A		B: 440 P	~	12
Osprey Pandion haliaetus	MB PV	A		B: 150 P	+++	8
Kestrel Falco tinnunculus	RB WV	А		B: 55,000 P	~	2
Merlin Falco columbarius	RB WV	А		B: 1,300 P	/++	12
Hobby Falco subbuteo	MB	G		B: 2,200 P	+++	
Peregrine Falcon Falco peregrinus	RB WV	Α		B: 1,400 P	+++	8
Red Grouse Lagopus lagopus	RB	А		B: 160,000 P		7
Ptarmigan Lagopus muta	RB	G		B: 10,000 P	?~	
Black Grouse Tetrao tetrix	RB	R		B: 5,100 M		12
* Capercaillie Tetrao urogallus	RB	R	$1837^{\dagger}$	B: 1,200 I		10
* Red-legged Partridge <i>Alectoris rufa</i>	RB	-	1770	B: 140,000 T	~	2
Grey Partridge Perdix perdix	RB	R		B: 72,000 P		2
Quail Coturnix coturnix	MB	R		B: 150 M	-	-
Zum Commercommer	1110	11		D. 150 101		

* Pheasant Phasianus colchicus	RB	-		B: 1,700,000 F	++	2
<sup>*</sup> Golden Pheasant Chrysolophus pictus	RB	-	1870s	B: 100 P		
<sup>*</sup> Lady Amherst's Pheasant <i>Chrys. amherstiae</i>	RB	-	1890	B: 90 P		
Water Rail Rallus aquaticus	RB WV	Α		B: 700 P	?	
Spotted Crake Porzana porzana	MB	А		B: 70 M	+	8
Corncrake Crex crex	MB	R		B: 1,100 M	/+	8
Moorhen Gallinula chloropus	RB WV	G		B: 160,000 T	~	4
_				W: 750,000 I	?	
Coot Fulica atra	RB WV	G		B: 110,000 P	++	4
				W: 170,000 I	~	6
Crane Grus grus	CB	А	$1981^{\dagger}$	B: 5 P	+	
Oystercatcher Haematopus ostralegus	RB WV	А		B: 110,000 P	+++/~	4
, i i				W: 320,000 P	~	6
Black-winged Stilt Himantopus himantopus	CB	_	1945	B: <1 P		
Avocet <i>Recurvirostra avosetta</i>	RB WV	А	1941 <sup>†</sup>	B: 900 P	+++	8
			17.11	W: 3,400 I	~/+++	6
Stone-curlew Burhinus oedicnemus	MB	R		B: 310 P	/++	8
Lapwing Vanellus vanellus	RB WV	A		B: 150,000 P	++/	2
Lapwing vanenas vanenas		11		W: 1,800,000 I	++	6
Golden Plover Pluvialis apricaria	RB WV	G		B: 46,000 P	-	12
Golden Hover Truvians apricaria		U		W: 250,000 I	- ~/+++	6
Grey Plover Pluvialis squatarola	WV	А		W: 53,000 I W: 53,000 I	~/+++ +++/-	6
		A G	1029	,		
Little Ringed Plover Charadrius dubius	MB		1938	B: 950 P	+	6
Ringed Plover Charadrius hiaticula	RB WV	А		B: 8,400 P		12
	1 (D			W: 32,000 I	-	6
Dotterel Charadrius morinellus	MB	A		B: 630 M	?~	
Knot Calidris canutus	WV	A		W: 280,000 I	~	6
Sanderling Calidris alba	WV	G		W: 21,000 I	~	6
Little Stint Calidris minuta	PV	G		P: 450 I	~	
Temminck's Stint Calidris temminckii	CB PV	А	1934	P: 100 I	~	9
Curlew Sandpiper Calidris ferruginea	PV	G		P: 650 I		
Purple Sandpiper Calidris maritima	CB WV	А	1978	W: 18,000 I	~/	6
Dunlin Calidris alpina	MB WV	А		B: 9,500 P		
				W: 560,000 I		6
Ruff Philomachus pugnax	MB PV	А	1963 <sup>†</sup>	B: 40 M	~	8
				P: 700 I	-	6
Jack Snipe Lymnocryptes minimus	WV	G		W: 50,000 I	-/+	6
Snipe Gallinago gallinago	RB WV	Α		B: 52,000 P		1
				W: 100,000 I	~	6
Woodcock Scolopax rusticola	RB WV	А	1820s	B: 8,800 P		
Black-tailed Godwit Limosa limosa	MB WV	R	$1952^{\dagger}$	B: 50 P	~	8
				W: 15,000 I	+++	6
Bar-tailed Godwit Limosa lapponica	WV	А		W: 62,000 I	~	6
Whimbrel <i>Numenius phaeopus</i>	MB WV	A		B: 530 P	?-	
Curlew Numenius arquata	RB WV	A		B: 110,000 P	-	2
Spotted Redshank Tringa erythropus	WV	A		P: 500 I	?-	6
Redshank Tringa totanus	RB WV	A		B: 39,000 P		4
Redshank Tringa lotanus				W: 120,000 I	~	6
Greenshank Tringa nebularia	RB WV	G		B: 1,000 P	++	0
Greenshank Tringa nebutarta		U		P: 4,300 I	++	6
Graan Sandninger Tringa achronys	CB WV	А	1959	B: 2 P		0
Green Sandpiper Tringa ochropus		A	1939		+	
Wood Sondringer Twing a lange la	CD WW	٨	1050	P: 2000 I	+	0
Wood Sandpiper Tringa glareola	CB WV	A	1959	B: 10 P	++	8
Common Sandpiper Actitis hypoleucos	MB	G		B: 50,000 P	~/-	4
Turnstone Arenaria interpres	WV	A		W: 50,000 I	~	6
Red-necked Phalarope Phalaropus lobatus	MB	R		B: 40 M	~	8

Grey Phalarope Phalaropus fulicarius	PV	-		P: 300 I	~	9
Pomarine Skua Stercorarius pomarinus	PV	G		?		
Arctic Skua Stercorarius parasiticus	MB PV	G		B: 2,100 P	+++/-	11
Long-tailed Skua Stercorarius longicaudus	PV	G		?		
Great Skua Stercorarius skua	MB PV	А	1770s	B: 9,600 P	+++	11
Mediterranean Gull Larus melanocephalus	MB WV	А	1968	B: 110 P	~/+++	8
Little Gull Hydrocoloeus minutus	CB PV	G	1975	?	++	
Sabine's Gull Larus sabini	PV	-		P: 150 I	~	9
Black-headed Gull <i>Chroicocephalus ridibundus</i>	RB WV	А		B: 128,000 P	~	11
				W: 2,200,000 I	+++/-	14
Common Gull Larus canus	RB WV	А		B: 48,000 P	++	11
				W: 700,000 I	+++	14
Lesser Black-backed Gull Larus fuscus	MB WV	А		B: 110,000 P	++	11
				W: 130,000 I	+++	14
Herring Gull Larus argentatus	RB WV	А		B: 130,000 P		11
				W: 730,000 I	~	14
Yellow-legged Gull Larus michahellis	CB PV	-	1995	B: <1 P	+	
Iceland Gull Larus glaucoides	WV	G	1770	?	· ?+	
Glaucous Gull Larus hyperboreus	WV	G		· ?	?-	
Great Black-backed Gull Larus marinus	RB WV	G		B: 17,000 P	~	11
Great Brack Sucked Guil Laras marmus		0		W: 76,000 I	?++	14
Kittiwake Rissa tridactyla	RB WV	А		B: 370,000 P	~	11
Sandwich Tern Sterna sandvicensis	MB	A		B: 11,000 P	~	11
Roseate Tern Sterna dougallii	MB	R		B: 100 P		11
Common Tern Sterna hirundo	MB	G		B: 10,000 P	~	11
Arctic Tern Sterna paradisaea	MB	A		B: 53,000 P	+/-	11
Little Tern Sternula albifrons	MB	A		B: 2,000 P	+/-	11
Black Tern Chlidonias niger	CB PV	G		2,000 I ?	?	11
Guillemot Uria aalge	RB WV	Ă		B: 1,300,000 I	+++	11
Razorbill Alca torda	RB WV	A		B: 160,000 I	++	11
Black Guillemot Cepphus grylle	RB	A		B: 38,000 I	~	11
Little Auk <i>Alle alle</i>	WV	G		D. 50,000 I ?	?	11
Puffin <i>Fratercula arctica</i>	RB WV	A		B: 580,000 P	+	11
Feral Pigeon Columba livia	RB	G		B: 1,300,000 I	?	11
Stock Dove <i>Columba oenas</i>	RB WV	A		B: 310,000 T	+++	2
Woodpigeon Columba palumbus	RB WV	G		B: 2,700,000 T	+++	$\frac{1}{2}$
Collared Dove <i>Streptopelia decaocto</i>	RB	G	1955	B: 800,000 T	+++	$\frac{1}{2}$
Turtle Dove <i>Streptopelia turtur</i>	MB	R	1700	B: 44,000 T		2
* Rose-ringed Parakeet <i>Psittacula krameri</i>	RB	-	1971	B: 4,300 I	+++	2
Cuckoo <i>Cuculus canorus</i>	MB	А	1771	B: 14,000 P	~/	2
Barn Owl Tyto alba	RB	A		B: 4,000 P	, /~	12
*Little Owl Athene noctua	RB	-	1879	B: 8,700 P	?-	2
Tawny Owl Strix aluco	RB	G	10//	B: 190,000 P	~	2
Long-eared Owl Asio otus	RB WV	G		B: 2,400 P	?-	2
Short-eared Owl Asio flammeus	RB WV	A		B: 2,300 P		
Snowy Owl Bubo scandiacus	CB	-	1967	B: <1 P		
Eagle Owl <i>Bubo bubo</i>	RB	-	1984	B: 1 P		
Nightjar Caprimulgus europaeus	MB	R	1701	B: 4,600 M	/+	12
Swift Apus apus	MB	G		B: 110,000 P	?/-	3
Kingfisher Alcedo atthis	RB	A		B: 5,700 P	~	4
Bee-eater Merops apiaster	CB PV	-	1955	P: 20 I	++	9
Hoopoe Upupa epops	CB PV CB PV	G	1755	P: 120 I	~	9
Wryneck Jynx torquilla	CB PV CB PV	R		B: <1 P		,
	CDIV	11		P: 300 I	~	9
Green Woodpecker Picus viridis	RB	А		B: 24,000 P	+++	2
Great Spotted Woodpecker Dendrocopos major	RB WV	G		B: 41,000 P	+++	$\frac{2}{2}$
2		0		2. 11,000 1		2

Lesser Spotted Woodpecker Dendrocopos minor	RB	R		B: 2,200 P	~/	1
Woodlark Lullula arborea	MB	R		B: 3,100 P	+++	12
Skylark Alauda arvensis	RB WV	R		B: 1,700,000 T		2
Shorelark Eremophila alpestris	CB WV	-	1977	W: 300 I	?	
Sand Martin Riparia riparia	MB	А		B: 70,000 P	~	4
Swallow Hirundo rustica	MB	А		B: 780,000 T	~	2
House Martin Delichon urbicum	MB	А		B: 420,000 P		2
Tree Pipit Anthus trivialis	MB	А		B: 75,000 T	~/	2
Meadow Pipit Anthus pratensis	RB WV	А		B: 1,600,000 T		2
Rock Pipit Anthus petrosus	RB WV	G		B: 34,000 P	?-	
Water Pipit Anthus spinoletta	WV	G		W: <100 I	+	
Yellow Wagtail Motacilla flava	MB	А		B: 19,000 T		2
Grey Wagtail Motacilla cinerea	RB	А		B: 60,000 P	~	4
Pied Wagtail Motacilla alba	RB PV	G		B: 470,000 T	++/~	2
Waxwing Bombycilla garrulus	WV	Ğ		W: 100 I	?	
Dipper Cinclus cinclus	RB	G		B: 13,000 P	~	4
Wren Troglodytes troglodytes	RB	G		B: 8,000,000 T	++	2
Dunnock Prunella modularis	RB WV	Ā		B: 1,800,000 T	-/~	2
Robin Erithacus rubecula	RB WV	G		B: 5,500,000 T	, ~/++	2
Nightingale Luscinia megarhynchos	MB	Ă		B: 6,700 M	~	2
Bluethroat <i>Luscinia svecica</i>	CB PV	A	1968	P: 80 I	?	9
Black Redstart <i>Phoenicurus ochruros</i>	RB WV	A	1923	B: 50 P	++/-	8
Redstart <i>Phoenicurus phoenicurus</i>	MB	A	1725	B: 100,000 P	~	10
Whinchat Saxicola rubetra	MB	G		B: 21,000 P		3
Stonechat Saxicola torquatus	RB	A		B: 15,000 P	/+	3
Wheatear <i>Oenanthe oenanthe</i>	MB	G		B: 550,000 I	-	3
Ring Ouzel Turdus torquatus	MB	R		B: 6,900 P		12
Blackbird <i>Turdus merula</i>	RB WV	G		B: 4,800,000 T	-	2
Fieldfare <i>Turdus pilaris</i>	CB WV	A	1967	W: 680,000 I	?~	5
Song Thrush Turdus philomelos	RB WV	R	1707	B: 1,000,000 T		2
Redwing <i>Turdus iliacus</i>	RB WV	A	1932	B: 60 P	~	8
Redwing Turaus indeus		л	1952	W: 650,000 I	~ ?~	5
Mistle Thrush Turdus viscivorus	RB WV	А		B: 250,000 T	.~ -	2
Cetti's Warbler <i>Cettia cetti</i>	RB	G	1972	B: 650 M	-+++	8
Grasshopper Warbler Locustella naevia	MB	R	1972	B: 11,000 P	+++ /~	3
Savi's Warbler Locustella luscinioides	CB	R	1960 <sup>†</sup>	B: <10 P	/~ ++/	8
Aquatic Warbler Acrocephalus paludicola	PV	R	1900	P: 30 I?	++/ ~	9
Sedge Warbler Acrocephalus schoenobaenus	MB	G		B: 300,000 T	~	4
Marsh Warbler Acrocephalus palustris	MB	R		B: 30 P	~	8
					 ?+	
Reed Warbler Acrocephalus scirpaceus	MB CB PV	G G	1992	B: 91,000 P P: 70 I		2 9
Icterine Warbler <i>Hippolais icterina</i> Dartford Warbler <i>Sylvia undata</i>	RB	A	1992	B: 3,200 P	~	12
-	PV	A		P: 200 I	+++	9
Barred Warbler Sylvia nisoria		G			+~~~	
Lesser Whitethroat Sylvia curruca	MB	G		B: 64,000 T		2
Whitethroat Sylvia communis	MB	G		B: 930,000 T	/++	2
Garden Warbler Sylvia borin	MB MD WW	G		B: 190,000 T	~	2
Blackcap <i>Sylvia atricapilla</i>	MB WV	G		B: 920,000 T	+++	2
Pallas' Warbler <i>Phylloscopus proregulus</i>	PV	-		P: 110 I	+++	9
Yellow-browed Warbler <i>Phyll. inornatus</i>	PV	-		P: 430 I	+++	9
Wood Warbler <i>Phylloscopus sibilatrix</i>	MB	A		B: 17,200 M	~/	3
Chiffchaff Phylloscopus collybita	MB WV	G		B: 750,000 T	+	2
Willow Warbler <i>Phylloscopus trochilus</i>	MB	A		B: 2,000,000 T	~/	2
Goldcrest Regulus regulus	RB WV	A	10/2	B: 770,000 T	~	2
Firecrest Regulus ignicapilla	RB WV	A	1962	B: 170 M	++	8
Spotted Flycatcher Muscicapa striata	MB	R		B: 128,000 T		2
Red-breasted Flycatcher Ficedula parva	PV	-		P: 100 I	+	9

	MD	C		D 20 000 D	. 1	2
Pied Flycatcher <i>Ficedula hypoleuca</i>	MB	G		B: 38,000 P	+/-	3
Bearded Tit Panurus biarmicus	RB	A		B: 500 P	+++/~	10
Long-tailed Tit Aegithalos caudatus	RB	G		B: 260,000 T	++	2
Marsh Tit <i>Poecile palustris</i>	RB	R		B: 53,000 T		2
Willow Tit Poecile montana	RB	R		B: 8,500 T	 ?+	2
Crested Tit Lophophanes cristatus	RB	G		B: 2,400 P		10
Coal Tit <i>Periparus ater</i>	RB	G		B: 600,000 T	++/~	2
Blue Tit Cyanistes caeruleus	RB	G		B: 3,300,000 T	+	2
Great Tit Parus major	RB	G		B: 2,000,000 T	++	2
Nuthatch Sitta europaea	RB	G		B: 140,000 T	+++	2
Treecreeper Certhia familiaris	RB	G		B: 200,000 T	~	2
Golden Oriole Oriolus oriolus	MB	A		B: <10 P	++/	8
Red-backed Shrike Lanius collurio	CB	R		B: <1 P		8
	33737			P: 200 I	-	9
Great Grey Shrike Lanius excubitor	WV	-		W: 100 I	-	9
Jay Garrulus glandarius	RB WV	G		B: 160,000 T	~,	2
Magpie <i>Pica pica</i>	RB	G		B: 600,000 T	++/~	2
Chough Pyrrhocorax pyrrhocorax	RB	A		B: 450 P	++	12
Jackdaw Corvus monedula	RB WV	G		B: 1,000,000 T	++	2
Rook Corvus frugilegus	RB WV	G		B: 1,200,000 P	+	3
Carrion Crow Corvus corone	RB	G		B: 790,000 T	+++	2
Hooded Crow Corvus cornix	RB WV	-		B: 160,000 T	-	
Raven Corvus corax	RB	G		B: 12,000 P	?~	3
Starling Sturnus vulgaris	RB WV	R		B: 3,000,000 P		2
House Sparrow Passer domesticus	RB	R		B: 4,700,000 P		2
Tree Sparrow Passer montanus	RB	R		B: 68,000 T		2
Chaffinch Fringilla coelebs	RB WV	G		B: 5,600,000 T	+	2
Brambling Fringilla montifringilla	CB WV	G	1920	W: 920,000 I	?+	5
Serin Serinus serinus	CB PV	A	1967	P: 50 I	+++	9
Greenfinch Carduelis chloris	RB WV	G		B: 1,800,000 T	+	2
Goldfinch Carduelis carduelis	RB	G		B: 900,000 T	~	2
Siskin Carduelis spinus	RB WV	G		B: 360,000 T	++/?~	3
Linnet Carduelis cannabina	RB WV	R		B: 540,000 T		2
Twite Carduelis flavirostris	RB WV	R		B: 10,000 P		12
Lesser Redpoll Carduelis cabaret	RB	A		B: 25,000 P		2
Common Redpoll Carduelis flammea	CB WV	G		?		
Crossbill Loxia curvirostra	RB WV	G		B: 10,000 P	~	
Scottish Crossbill Loxia scotica	RB	R	1001	B: 780 P	?+	
Parrot Crossbill Loxia pytyopsittacus	RB	Α	1984	B: 30 P	?	
Scarlet Rosefinch Carpodacus erythrinus	CB PV	Α	1982	P: 150 I	+++	9
Bullfinch Pyrrhula pyrrhula	RB	R		B: 160,000 T		2
Hawfinch Coccothraustes coccothraustes	RB	A		B: 4,800 P	~/	12
Lapland Bunting Calcarius lapponicus	CB WV	G	1977	W: 350 I	~	
Snow Bunting Plectrophenax nivalis	RB WV	А		B: 90 P	++/-	8
		_		W: 11,000 I	-	
Yellowhammer Emberiza citrinella	RB WV	R		B: 790,000 T		2
Cirl Bunting Emberiza cirlus	RB	R		B: 710 P	/++	10
Ortolan Bunting Emberiza hortulana	PV	-		P: 60 I	~	9
Little Bunting Emberiza pusilla	PV	-		P: 30 I	++	9
Reed Bunting Emberiza schoeniclus	RB WV	R		B: 190,000 T	+/	2
Corn Bunting Emberiza calandra	RB	R		B: 10,000 T		2

<sup>†</sup> Had bred in historical times

<sup>1</sup> RB: Resident Breeder, MB: Migrant Breeder, CB: Casual Breeder, PV: Passage Visitor, SV: Scarce Visitor, WV: Winter Visitor

<sup>2</sup> I: individuals, P: pairs, F: Females, M: males, N: nests, T: territories, counts relate to breeding (B), passage (P) or winter (W) seasons.

<sup>3</sup> Source for trend information: 1 - Common Birds Census; 2 - Common Birds Census/Breeding Bird Survey; 3 – Breeding Bird Survey/Gibbons et al. (1993); 4 – Waterways Bird Census; 5 – Garden Bird Feeding Survey (Chamberlain et al. 2006); 6 –Wetland Bird Survey; 7 - Game and Wildlife Conservancy Trust Game Bag Counts

(Tapper et al. 1992); 8 – Rare Breeding Birds Panel; 9 – Fraser et al. (2007); 10 – RSPB data; 11 – Mitchell et al. (2004); 12 - Species specific surveys; 13 – British Birds Rarities Committee; 14 – Winter Gull Survey (Banks et al. 2009).